# FISHERIES BIOECOLOGY AT THE KHONE FALLS 

 (MEKONG RIVER, SOUTHERN LAOS)

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## FISHERIES BIOECOLOGY AT THE KHONE FALLS

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Figure 1:The Khone Falls area, including Hang Khone Village, Southern Laos Map prepared by Ole Heggen, Geography Department, University of
Victoria, Canada. This map is a generalized illustration only. The representation of political boundaries is not to be used for reference purposes.

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## EXECUTIVE SUMMARY

## GENERAL DESCRIPTION OF THE FISHERY

- The database that this report is based upon covers six years of monitoring of artisanal fisheries (over 20 nuclear fishing families) in Hang Sadam and especially Hang Khone villages, just below the Khone Falls in Khong District, Champasak Province, Southern Laos.
- Forty-one fishing methods were monitored. Gill nets with various mesh sizes were the most frequently used gear; they make up to $73 \%$ of the number of fishing operations included in the database. A number of categories of gill nets, based on mesh sizes (from 2.5 to 30 cm ), were monitored throughout the six years, and seven fishing methods were monitored consecutively for at least five years.
- There are two main peak periods in fishing activity: from November to March (most intense period), and in June. Fishing intensity is the least between August and October. This trend is driven by the frequent utilization of gill nets during these months.
- Four main groups of fishing gears can be identified: gears operating all year long; gears targeting migrating fish at the beginning of the rainy season (June-July); gears operating at the end of the rainy season (October to January) and gears operating in the dry season (January to May).


## INTERANNUAL TRENDS: ISSUES AND CONSTRAINTS

Given the very particular context of these fisheries, which target more than one hundred species of fish, some migratory and others more or less sedentary, with a diversity of gears, an analysis of trends in annual overall catches would face multiple constraints and biases. We detail this point through taking the Probarbus jullieni fishery as an example, and show that such contexts challenge classical fishery science approaches.

## COMPARISON OF CAMBODIAN BAG-NET (DAI) FISHERY AND LAO FENCE-FILTER TRAP (TONE) FISHERY

It was not possible to draw significant conclusions from an extensive analysis of available data about the correlation between the catch of the tone fishery in the Khone Falls area and the dai fishery in Cambodia. This was due to a limited number of points ( $\mathrm{n}=5$ years).

## ABUNDANCE PATTERNS OF 110 SPECIES

The abundance patterns of 110 fish taxa (i.e. fish identified at the species level or at least at the genus level) in catches have been analysed; they often indicate migration patterns. Three major groups of fish have been identified:
i) species present during the dry season (peak in January), some of them exhibiting a small secondary peak in catches at the beginning of the rainy season (May-June);
ii) species present during two equivalent periods (dry and wet season respectively) or being regularly distributed all year long (no evident migrations); and
iii) species showing a dominant or exclusive abundance at the beginning of the rainy season (May-June).
Among the 110 taxa studied, 90 exhibit strong patterns of sudden abundance in catches. Most of these species are migrating.

## MIGRATION TRIGGER OF PANGASIUS KREMPFI

The catfish Pangasius krempfi provides a clear example of a migration triggered by a water level rise. The migration of this species occurs suddenly at the beginning of the wet season, and lasts in general 40-50 days. There are generally about a dozen days of intense migration peak each year. This points to the likely consequences of building dams that regulate the hydrology of the Mekong River and its tributaries on $P$. krempfi and on other species responding similarly to hydrological triggers.

## DOMINANT SPECIES: LIFE HISTORY KEY FACTS, MIGRATIONS AND MIGRATION TRIGGERS

Dominant taxa in catches
There are forty-seven taxa for which at least 50 individuals have been caught on average per year over six years

## Length of fish caught

- $61 \%$ of dominant taxa caught have a maximal size greater than 25 cm , but
$-85 \%$ of fishes actually caught (in terms of biomass) belong to taxa whose maximal size is less than 25 cm . This illustrates the abundance of small species in the catch, despite the diversity of a fish community in which large species are largely represented.


## Species migrations

Ninety-six percent of dominant taxa are not present all year long, and thus undertake migrations or become impossible to catch at certain periods of time. This figure highlights the importance of migration behavior among dominant species of the Khone Falls fisheries.

## Species, discharge and migration triggers

The highest biodiversity appears in catches for the lowest discharge levels. This underscores the importance of dry season water levels for fish and fishers, as well as that of rising waters at the end of the dry season.

1) at least $55 \%$ of taxa are sensitive to discharge among the dominant taxa of the Khone Falls.
2) Ninety six percent of the fish caught (expressed in terms of biomass) are highly sensitive to discharge in the Khone Falls fisheries.
Given this extremely high sensitivity of fishes to discharge and discharge variations, when flows are regularized, a very significant share of the fishery resource will be permanently impacted and might disappear from the catch.

## DEEP POOLS AS DRY SEASON REFUGES

Catch-per-unit-effort (CPUE) of the gill nets set in deep-water pools in February and March is three to twelve times higher than CPUE of gill nets used in surface during the same period, and with the same mesh sizes.

Among the 30 dominant species caught during this monitoring, only two showed a preference for surface waters. Ten species showed a clear preference for deep-water pools in the dry season; they consisted of seven siluriforms (catfishes) and three cypriniforms (carps).

The average weights of 18 species caught at least five times in each environment were analysed. For ten species, the individuals caught in deep pools are 27 to $78 \%$ larger than those caught near the surface; for two species the individuals caught at the surface are larger than those caught in deep-water pools.

## INTRODUCTION

This study results from a collaboration between the Global Association for People and the Environment (GAPE), which provided the data and the field knowledge used in this report, and the WorldFish Center (formerly ICLARM), which analysed data and developed an interactive interface for data display.

The data were collected as part of the Lao Community Fisheries and Dolphin Protection Project (LCFDPP) and the Environmental Protection and Community Development in Siphandone Wetlands Project (EPCDSWP), with the latter project being implemented by the Italian non-governmental organization CESVI Cooperation and Development, and funded by the European Union. At the WorldFish Center, data analysis and publication result from a project entitled "Conservation of Aquatic Biodiversity / Mekong Initiatives" also funded by the European Union.

In this study, over 20 fishers or fishing families using multiple seasonal gears on the southern end of Khone Island and the adjacent southern


Figure 2: Indochina, Mekong River Basin and Khone falls

end of Sadam Island (mainstream Mekong River, Southern Lao PDR) were monitored between 1993 and 1999. Both locations are just below the Khone Falls fault line, and just north of the border with Cambodia.

The use of the artisanal fishing gears in the Khone Falls are described, new information about the possible migration patterns of 110 Mekong fish taxa is provided, hydrological triggers of fish migrations are detailed, and the crucial role of deep-water pools in the mainstream Mekong River as refuges for fish in the dry season is demonstrated.

The overall objective of this study is to contribute to biodiversity conservation and the sustainable management of inland fisheries in the Khone Falls, a zone of major ecological and livelihood importance in Laos and in the Mekong River Basin.

## F FISHING GEARS, FISHING OPERATIONS AND FISH CAUGHT

The monitoring of the Khone Falls fisheries started on 4th March 1993, and continued until 30th May 1999. This represents monitoring of six full years and five full seasons of fishing (a fishing season is traditionally considered to start in October). Over 20 fishers and their immediate families were monitored in Hang Sadam Village and especially Hang Khone Village, along with a few others whose catches were only monitored periodically.

## During these years:

- Forty-one gears were initially monitored; they were lumped for analyses into 32 categories (due to overlapping mesh sizes among initial gill net categories; Table 1);
- 20,222 fishing operations were analyzed (Table 2 and Figure 3);
- 138 species or taxa were caught;
- 666,000 individual fishes plus many bulk weighed fishes were caught. Overall, the database includes a biomass of 53 metric tons (Table 3).


Table 1: Names of the gears monitored

| Lao gear |
| :--- | :--- |
| name |$\quad$| Gear <br> description |
| :--- |
| Bet khen siap <br> mengkhinai |
| Set hook baited <br> with mole cricket |
| Bet teuk siap <br> khikadeuan | | Pole and line |
| :--- |
| baited with worm |$|$| Bet teuk siap <br> koung | Pole and line <br> baited with shrimp |
| :--- | :--- |
| Chan | Falling-door trap |
| Chap pa kap | Fish caught <br> with hands |
| Chip | Cylindrical <br> current trap |
| He po | Large meshed <br> cast-net |
| He soi | Small meshed <br> cast-net |
| Kha | Branch bundle fish <br> attractant <br> pull basket trap |
| Lai tao siap | Floating hook <br> baited with Ficus sp. <br> fruit |
| mak deua |  |
| Lai tao siap | Floating hook <br> baited with Crayratia <br> trifolia fruit |


| Lao gear <br> name | Gear <br> description |
| :--- | :--- |
| Lan | Upright basket trap <br> baited with bran |
| Li | Khone Falls <br> wing trap <br> (monsoon season) |
| Lope gnai | Large funnel <br> basket trap |
| Lope none | Hang Sadam Village <br> funnel basket trap |
| Lope tang | Hang Sadam Village <br> standing funnel <br> basket trap |
| Mong <br> $\mathbf{1 2 - 1 8}$ | Set gill net <br> $12-18 \mathrm{~cm}$ |
| Mong <br> $\mathbf{1 8}-\mathbf{3 0}$ | Set gill net <br> $18-30 \mathrm{~cm}$ |
| Mong <br> $\mathbf{2 . 5}$ | Set gill net <br> $2.5-3 \mathrm{~cm}$ |
| Mong <br> $\mathbf{4 - 9}$ | Set gill net <br> $\mathbf{4 - 1 1} \mathrm{cm}$ |
| Mong yone | Deep set gill net <br> $12-18 \mathrm{~cm}$ |
| $\mathbf{1 2 - 1 8}$ | Mong yone <br> $\mathbf{4 - 9}$ |
| $\mathbf{4 - 9 ~ c m ~ s e t ~ g i l l ~ n e t ~}$ |  |


| Lao gear name | Gear description |
| :---: | :---: |
| Phiak siap khikadeuan | Longline baited with worm |
| Phiak siap mak deua | Longline baited with Ficus $s p$. Fruit |
| Phiak siap mak houn | Longline baited with Crayratia trifolia fruit |
| Phiak siap mengkhinai | Longline baited with mole cricket |
| Phiak siap pako | Longline baited with Gyrinocheilus pennocki fish |
| Phiak siap pasoi | Longline baited with cyprinid fish |
| Souang | Wedge cone trap |
| Tone | Khone Falls fence filter trap (dry season) |
| Tone houay | Hang Sadam Village stream fence filter trap |
| Tone na | Hang Sadam Village rice field fence filter trap |
|  |  |

Table 2: Fishermen and their total effort in the database

| Fisher code | 1 | 3 | 4 | 2 | 5 | 10 | 6 | 7 | 8 | 9 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fishers names | Vilay;Bounpheng; <br> Sounthone; <br> Nouphai (family) | Songma | Sai | Keo | Sout | Bounheng | Mai | Tha | Sit |  |
| \# of fishing <br> operations | 3556 | 3196 | 1975 | 1871 | 1408 | 1329 | 1236 | 1011 | 849 | 704 |


| Fisher code | 13 | 22 | 11 | 15 | 12 | 16 | 23 | 20 | 14 | 17 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \# of fishing operations | 308 | 261 | 255 | 152 | 102 | 92 | 40 | 36 | 23 | 22 |


| Fisher code | 18 | 21 | 30 | 24 | 19 | 25 | 26 | 40 | 50 | 0 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \# of fishing operations | 22 | 20 | 10 | 6 | 5 | 5 | 5 | 5 | 5 | 1713 |



Fisher code
Figure 3: Fishers and their fishing effort during 6 years

Table 3: Abundance and biomass of fish caught

| Fishing season |  |  | Fishing season |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1992-1993 | Number | 1846 | 1996-1997 | Number | 600737 |
|  | Biomass (kg) | 9527 |  | Biomass (kg) | 6668 |
| 1993-1994 | Number | 15612 | 1997-1998 | Number | 18640 |
|  | Biomass (kg) | 6956 |  | Biomass (kg) | 6870 |
| 1994-1995 | Number | 18403 | 1998-1999 | Number | 3070 |
|  | Biomass (kg) | 8358 |  | Biomass (kg) | 5318 |
| 1995-1996 | Number | 8393 | Total Number | 666701 |  |
|  | Biomass (kg) | 9757 | Total biomass (kg) | 53455 |  |

Note: In 1992-1993 and 1998-1999, only a fraction of the October-September fishing season was monitored; the fishing seasons fully monitored are highlighted above in italics.


## 3 TEMPORAL USE OF KHONE FALLS FISHING GEARS

## 3-1) Annual use of fishing gears

Gill nets were the most frequently sampled fishing gears; they made up to $73 \%$ of fishing observations (data and results are detailed in a companion technical report). Three categories of gill nets were mainly used and monitored throughout the six years: gill nets $4-11 \mathrm{~cm}$; gill nets $12-16 \mathrm{~cm}$; gill nets $18-30 \mathrm{~cm}$. However, $2.5-3 \mathrm{~cm}$ meshed gill nets gained popularity during the latter years of the study, but were not intensively monitored over a long period, although some observations of catches were recorded.

Thirteen fishing methods have been monitored less than 70 times in seven years (arbitrary value representing less than $0.3 \%$ of the total fishing effort). The general use of the fishing gears or methods is illustrated in Figure 4.

It should be noted that these results reflect the gears monitored, and not the exact number of gears operating in the area. For example, 2.5 cm meshed gill nets are underrepresented; only limited numbers of small meshed gill nets were initially used and monitored, but their use by fishers rose dramatically over the course of monitoring, which is not illustrated well by the data.

Tone traps (fence-filter traps operated during the dry season) and $l i$ traps (wing traps operated during the monsoon season) are also quite underrepresented because, for logistical reasons, i) only one fishing site for each of the two gears was monitored, and ii) fishers who were sampled for other gears were not sampled for their $l i$ and tone traps, even though many had them. Thus tone and $l i$ traps are some of the most important gears in the Khone Falls area, but the data does not reflect this reality well.

Generally, the database is most useful for i) considering certain fisheries by themselves, and over a period of time; and ii) considering the presence and abundance of fish species at certain times of the year.



Figure 4: Annual number of operation for each fishing method monitored

## 3-2) Monthly use of fishing gears

It is not possible, due to sampling biases and other constraints, to use the database to determine the monthly variability in the use of fishing gears. For instance $l i$ traps are very active in May, but due to logistical constraints, only one set of these traps was monitored. Thus this database results from the monitoring during six years of 21 nuclear fishing families, but is not exactly representative of the overall fishing activity in the area.

In order to give an idea of the seasonal variation of fishing activity in the whole area, we propose below a qualitative estimate based on the extensive field experience of the second author. The monthly intensity of fishing for the 19 gears that represent $99.7 \%$ of the total fishing effort has been ranked according to the following scale:
0 : gear not used or not set
1: few gears set or little catch
2: moderate fishing
3: intense fishing

The result is given in Table 4 and the global temporal pattern in Figure 5.


Table 4: Relative intensity of fishing per gear and per month at Ban Hang Khone
0 : gear not used or not set $\quad 1$ : few gears set or little catch
2: moderate fishing
3: intense fishing

| Gear name | Oct | Nov | Dec | Jan | Feb | Mar | Apr | May | Jun | Jul | Aug | Sep |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Gill net_set 12-16 | 1 | 2 | 2 | 2 | 1 | 1 | 1 | 3 | 3 | 2 | 1 | 1 |
| Gill net_set 18-30 | 1 | 3 | 3 | 2 | 1 | 1 | 1 | 2 | 2 | 1 | 1 | 1 |
| Gill net_set 4-11 | 1 | 2 | 3 | 3 | 2 | 2 | 1 | 1 | 2 | 1 | 1 | 1 |
| Longline baited_cyprinid fish | 0 | 1 | 0 | 2 | 2 | 2 | 1 | 1 | 3 | 3 | 1 | 0 |
| Trap_pull basket and branch bundle | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 3 | 1 | 0 |
| Trap_falling door | 2 | 3 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 |
| Trap_wing K. Falls | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 3 | 2 | 0 | 0 |
| Trap_fence filter K. Falls | 0 | 0 | 1 | 3 | 3 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| Gill net_set 2.5-3 | 1 | 1 | 1 | 3 | 2 | 2 | 2 | 3 | 2 | 1 | 1 | 1 |
| Longline baited_mole cricket | 2 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 |
| Trap_funnel basket H. Sadam V. | 1 | 3 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| Trap_current cylindrical | 0 | 2 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gill net_deep set 4-9 | 0 | 0 | 0 | 1 | 2 | 3 | 2 | 0 | 0 | 0 | 0 | 0 |
| long line baited_worm | 2 | 3 | 1 | 1 | 1 | 0 | 0 | 0 | 1 | 1 | 1 | 1 |
| long line baited_G. pennocki | 0 | 0 | 0 | 1 | 2 | 3 | 3 | 1 | 0 | 0 | 0 | 0 |
| long line baited_Ficus fruit | 0 | 0 | 0 | 1 | 3 | 3 | 2 | 1 | 0 | 0 | 0 | 1 |
| cast net sm. Meshed | 1 | 1 | 1 | 2 | 3 | 3 | 3 | 2 | 1 | 1 | 1 | 1 |
| Trap_wedge cone | 0 | 1 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gill net_deep set 12-18 | 0 | 0 | 1 | 1 | 3 | 3 | 1 | 0 | 0 | 0 | 0 | 0 |

Overall, the months with the most intense fishing are November to March, whereas August to September is the period with the least fishing effort and catches.

Among the gears, four main groups of fishing gears can be identified (Figure 6):

- Gears operated mostly at the beginning of the dry season, between October and January; these gears are longlines baited with mole cricket; Hang Sadam Village funnel basket traps; cylindrical current traps; falling-door traps; and wedge cone traps;


Figure 5: Monthly average intensity of fishing and discharge in Khone Falls area

- Gears targeting migrating fish in June and July (in particular, Khone Falls wing traps and branch bundle fish attractant pull basket traps);
- Gears operating all year long; these gears are gill nets of various sizes; and long lines baited with Cyprinid fish and cast nets;
- Gears operating in the dry season, mostly from January to May; these are Khone Falls fence- filter traps, deep-water gill nets; and long lines baited with Ficus fruit or Gyrinocheilus pennocki fish.


Figure 6: Khone Falls main fishing gears and their use in time


## 4 interanual trenos: ISSUES AND CONSTRAINTS

## 4-1) Interannual trends in overall catches

Given the availability of catch data over six years, it is tempting to study the interannual variation of fish catches at the Khone Falls. However, analysing trends in overall catches over several years by lumping the catches of all gears would not be relevant for several reasons:

1) A good (hydrological) year for species $A$ can be a bad one for species $B$, and given the species diversity in the fishery ( 138 species caught), lumping them all together would only blur the specific trends.
2) The annual abundance of some fish species is heavily influenced by the hydrological regime during the same year (e.g. small opportunists like Henicorhynchus spp.; Deap Loeung 1999, Baran et al. 2001a), while that of long-lived species (e.g. large catfishes) is buffered over several years. The abundance of such long-lived fishes seems to be declining, while changes in the catches of small opportunists are less evident (Van Zalinge \& Nao Thouk 1999, Sverdrup-Jensen 2002, Baird 2005d). Therefore, the response of catches to environmental factors over six years might be blurred by changes in the structure of fish communities due to fishing itself.
3) Following previous remarks, interannual trends in abundance should be analysed for coherent groups of fish only (e.g. black/white fishes at least).
4) Given i) the very high selectivity of the different fishing gears; ii) the specific choice of fishing methods depending on the hydrology and on the corresponding migrating species, and iii) the shift from one target species to another for one given gear at different times of the year. Thus, lumping all gears together would certainly blur the detailed trends, as shown in Figure 7:

5) Comparisons between years should, as pointed out above, be done on the basis of a standardized effort; however standardization (i.e. calculations based on the smallest temporal sequence is common) might force the removal of some months/weeks during which migration peaks occur. Fish abundance and catches during peak periods can be 100 times as much as catches during other weeks/months. Therefore, such standardization would create a major bias.
6) A quick overview of overall yearly trends (not detailed here) is not consistent with


Figure 7: Interannual trends in CPUE for 2 types of gears what fishers say about "good" and "bad" years. Fishers, for instance, mightconsider that a year is "good" when catches of a high-valued species like Probarbus are good, while catching the same biomass of small low value fish, such as Henicorhynchus, might not indicate "a good year".

Incidentally, these arguments outline the need for an in-depth and systematic analysis of the Khone Falls fisheries at the species level.

## 4-2) Interannual trends in catches of a migrating species

In the case of fisheries targeting migratory river species, classical approaches in fisheries science might be irrelevant due to very specific constraints and biases inherent to flooding rivers. The Probarbus jullieni fishery provides an example of such inadequacies.


Figure 8: Probarbus jullieni

Probarbus jullieni is a large Cyprinid found in the mainstreams of a number of large rivers in mainland Southeast Asia. In the Mekong River, it is believed to undertake short migrations (Singhanouvong et al. 1996a, Baird 2005d), although it has also been reported by some to conduct longer migrations (Pantulu 1986, Poulsen and Valbo-Jørgensen 2000). Observations of the species have resulted in the hypothesis that it might consist in a series discrete stocks that undertake migrations over relatively short distances (MFD 2002).

This excellent tasting and high-priced food fish can reach 150 cm and is mainly caught in large-meshed gill nets during its spawning season. Formerly abundant, it has been subject to serious long-term decline throughout its range and is now classified as endangered by IUCN (Baird 2005d). Its ecology and fishery has been extensively described in Baird (2005d). FishBase should also be consulted for details regarding this species (www.fishbase.org).

The database indicates the daily total catches of $P$. jullieni during six years at Hang Khone (Figure 9):

The data exhibit clear seasonal patterns, with sudden high abundance levels beginning in November-December every year. Over a period of years, data availability raises the following question: what is the long-term trend in catches of Probarbus jullieni? In other words, can a decline be identified in the catches of this endangered species?


Figure 9: temporal pattern of Probarbus jullieni catches

Standard as well as revised fishery science approaches (e.g. Holden and Raitt 1974, Caddy and Mahon 1995) would support standardizing these data and expressing them in terms of CPUE levels prior to making interannual comparisons. In the case of gill nets, CPUE is expressed in terms of kilograms of fish per hour per square meter of net, or, if the surface area is not exactly known, in terms of kilograms of fish per hours of fishing with a standard net.

Before addressing the above issue, we examined data at a smaller time scale, on a daily basis, and it became apparent that fishing seasons significantly vary in terms of their beginnings, ends and durations, even when daily catches remain in the same range (Figure 10; details of data for the 1997-1998 and 1998-1999 fishing seasons).


Figure 10: Comparison of 2 fishing seasons of Probarbus jullieni

The standardized comparison of CPUEs between 97-98 and 98-99 is included in Table 5. We have also integrated the number of days of catches (starting from the first day when Probarbus jullieni is caught in the season, to the last day).

Table 5: Catch of Probarbus jullieni, corresponding effort and CPUE

| Season |  | $93-94$ | $94-95$ | $95-96$ | $96-97$ | $97-98$ | $98-99$ |
| :---: | :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| A | Total catches (kg) | 509.21 | 552.65 | 552 | 139.3 | 247.8 | 154.3 |
| B | Total effort (hours of net) | 1476 | 3420 | 1728 | 612 | 1392 | 672 |
| A/B | CPUE (kg/hour of net) | 0.345 | 0.162 | 0.319 | 0.228 | 0.178 | 0.230 |
| C | Nb of days of peak abundance | 65 | 74 | 61 | 24 | 40 | 48 |
| (A/B)xC | CPUE integrating peak abundance duration | 22.4 | 12.0 | 19.5 | 5.5 | 7.1 | 11.0 |

The conclusions are that:

1) Standard CPUE analysis leads to the conclusion that fish were scarcer in 97-98 than in 96-97, as less fish were caught per hour of fishing.
2) The fishing period was much shorter in 96-97 than in 97-98. During years with long fishing seasons, fish can be caught at the same rate (i.e. same CPUE) but over a longer period of time, which results in larger catches overall.
3) Fishes were available much longer in 97-98 than in 96-97, and therefore overall catches were greater in 97-98. This illustrates a higher abundance of Probarbus jullieni in 97-98 than in 96-97.


Figure 11: Interannual trends as concluded from 2 different approaches


Figure 12: Interannual trend in the catch of Probarbus jullieni in Khone Falls
4) Conclusions drawn from standard CPUE analysis contradict conclusions that integrate peak catch duration data (Figure 11). However when tested versus facts (points 2 and 3 above), the approach based solely on CPUE is proven to be irrelevant, at least in this particular case
5) A possible way to partially overcome this problem is to multiply the CPUE by the number of days of fishing. This is possible under the assumption that fishers are skilled, and that they set their nets in space and time in a way that maximizes catches. This is the case of the Khone Falls fishers, who are particularly knowledgeable about the behavior of fishes (Roberts and Baird 1995, Baird et al. 1999a, Baird 2005a).

Thus the interannual trends in catches for the Probarbus jullieni fishery are illustrated in Table 5 and in Figure 12. These six years of monitoring show that the variability in catches is high. Despite the claims of fishers who probably focus on biomass caught, with less knowledge or consideration of variability in fishing effort, the data indicate that there was a decline in the catches of $P$. jullieni over time, but this trend is not very significant, with a regression coefficient of only 0.46 .

An important lesson from this analysis is that in terms of ecology and interannual trends, a standardized analysis focusing on CPUEs only leads to erroneous conclusions. Conversely an analysis focusing on biomass only would also be biased, as illustrated below.

| YearA |  | Catch |  |  | CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | x | $\bigcirc$ | $y$ | q | $x / y=z$ |
| A |  |  | 2 y | Q | $2 \mathrm{x} / 2 \mathrm{y}=\mathrm{z}$ |
|  |  |  |  |  |  |
| CATCH ANALYSIS bias: variable effort |  |  |  | CPUE ANALYSIS <br> bias: catch and effort correlation |  |
| (Catch is bigger in year B , but then fishing is also twice more intensive) |  |  |  | (CPUE remains the same from year to year, even when catch increases two fold) |  |

Figure 13: Possible biases in the analysis of interannual trends for migrant species

In the case of migrating river fishes harvested over a limited period of time every year, an assessment of production from year to year should result from a tri-dimensional approach encompassing CPUE, but also biomass and duration of the fishing season (Figure 14). In addition, as indicated by Baird (2005d), socioeconomic issues must also be considered. For example, gill nets used in the later years of the study were generally older and less efficient than in earlier years. This was because the fishery was believed to be in decline, and fishers were less willing or able to invest in new gill nets in the latter years. There were also reportedly more filamentous algae in the water in later years, reducing the efficiency of individual gill nets. Also, due to increased gill net theft, fishers were less willing to leave gill nets in the water during marginal fishing periods in the latter years as compared to the earlier years of the study.


Figure 14: Parameters of importance in the monitoring of migrating species production

## 5 <br> RELATIONSHIP BETWEEN A LAO AND A CAMBODIAN FISHERY

## COMPARISON OF THE FENCE - FILTER TRAP (TONE) AND BAG-NET (DAI) FISHERIES

The tone (or fence-filter trap) fishery in Hang Khone Village consists of a variety of different types of weirs and basket traps set in and around the Khone Falls area, the only part of the lower mainstream Mekong River where the geography allows for the use of these traps. The fishery operates from December to March, with peaks strongly associated with new moon periods (Baird et al. 2003). The species that are dominant in catches are all small highly migratory Cyprinids such as Henicorhynchus lobatus ${ }^{1}$, and Paralaubuca typus (51 and 33\% of catches respectively; Baird et al. (2003). This fishery was monitored during five seasons from 94-95 to 98-99 and is extensively described in Roberts and Baird (1995), Baird (1998) and especially in Baird et al. (2003).

The dai (or bag-net) fishery in the Tonle Sap River (Cambodia) operates from October to March; it mainly targets small migrating Cyprinids such as Henicorhynchus spp. and Paralaubuca typus (37 and $12 \%$ respectively).

This fishery has been monitored from 94-95 up to now and is described in Lieng et al. (1995), Deap Loeung (1999), Hap Navy and Ngor Peng Bun (2000), Ngor Peng Bun and Hem Chanthoeun (2000), and Baran et al. (2001). The catches of the dai fishery in Cambodia and the tone fishery in the Khone Falls area in Southern Laos were compared by Baird et al. (2003).

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The difference of scale between these two fisheries is important:

- Limited unit size in Laos, large gear in Cambodia.
- Annual average yields of about 3 tonnes per year for one fishing gear in Laos, and about 200 tonnes per gear in Cambodia.
- 1 unit monitored in Laos (of about 200 in use throughout the Khone Falls area), 63 units monitored in Cambodia (all the units in use).

Despite their difference in scale, these two fisheries are very similar in nature, as they both target the same species at around the same time of the year. Baird et al. (2003) reported that both fisheries probably target many of the same fish, with those fish taking about 20 days on average tomigrate upstream from where they are targeted by the dai fishery to where they are targeted by the tone fishery Henicorhynchus spp. and Paralaubuca typus, the
two dominant taxa in the catches of both fisheries, migrate between October and March from the Tonle Sap Great Lake (through the Tonle Sap River where the dai fishery operates) to the Khone Falls, where the tone fishery operates. When the water recedes in the Tonle Sap, Henicorhynchus spp. migrate down to the Mekong (Lieng et al. 1995) and from October to February the fish continue their journey up the Mekong, past the Khone Falls (Baird et al. 2003; Poulsen and Valbo-Jørgensen 2000).

Lao fishers operating the tone fishery believe that at this time of year a good fishing season in Cambodia


Figure 16: One dai unit in the Tonle Sap River corresponds (in relative terms) to a poor fishing season in Laos. They also believe that catches of the tone fishery in Laos reached historical record highs in 1975-1979, when the Khmer Rouge interrupted large-scale fishing operations. These claims have never been substantiated, and call for a more detailed comparison of the two fisheries. Baird et al. (2003) show that between 1995 and 1999 good dai fishing years tend to result in relatively poor catches in the tone fishery, but there are not enough years of data to significantly represent this negative correlation statistically.

An extensive analysis of available data has been undertaken by the authors in order to determine if the tone fishery catches were positively or negatively correlated to those of the dai fishery and to the hydrological level in each zone. However, the limited number of points ( $\mathrm{n}=5$ years) led to contradictory or spurious correlations (e.g. catch of year y with water level of year $y+1$ ). Subsequently, it was impossible to provide, based on existing monitoring data, clear-cut conclusions about the relationship between the two fisheries.

In both cases, catches are dominated by Henicorhynchus spp. and Paralaubuca typus (between 50 and $80 \%$ ), whose similar migrating behavior is well known; however 30 to $50 \%$ of catches are due to other species exhibiting various migratory patterns (see section on migrations), and these species might blur a trend noticed by fishers, and due to only a few species. It is therefore recommended that, beyond the scope of this exploratory report, comparisons between tone and dai fisheries should be deepened at the species level. The following section standardizes taxonomic categories for such an analysis.


Figure 17: October-March migrations of dominant species caught in Tone and Dai fisheries

The analysis of the tone and dai data sets highlighted several discrepancies between the fish taxonomy used in Southern Laos and in Cambodia. The intricacies of the taxonomy of Mekong fishes have already been pointed out elsewhere (Rainboth 1996, MRC Mekong Fish Database), and the resulting confusions have been highlighted (Baird et al. 1999a, Baran et al. 2001b, Baran and Chheng 2003).

In view of a detailed comparison of catches in the two fisheries for which different Latin names were used, it was necessary to adopt a common taxonomic system, detailed in Table 6.

Table 6: Taxonomic standardization for species caught in dai and tone fisheries

| Latin name (dai fishery database) | Lumped Latin name (Khone Falls fisheries database) | Latin name used | Reason |
| :---: | :---: | :---: | :---: |
| Acantopsis sp. | Acantopsis sp. or spp. | Acantopsis sp. | Species unknown OR uncertain identification |
| Achiroides leucorhynchos | Achiroides spp. | Achiroides spp. | Uncertain identification (2 close species) |
| Barbichthys thynnoides | Thynnichthys thynnoides | Thynnichthys thynnoides | Barbichthys thynnoides does not exist |
| Barbodes gonionotus | Hypsibarbus lagleri/ Hypsibarbus wetmorei | Barbodes/Hypsibarbus sp. or spp. | Dispute over identification |
| Botia sp. | Botia modesta/Botia spp | Botia spp. | Several possible species |
| Cyclocheilichthys spp. (non enoplos) | Cyclocheilichthys spp. non enoplos | Cyclocheilichthys spp. non enoplos |  |
| Henicorhynchus siamensis | Henicorhychus lobatus, H. siamensis and H. lineatus | Henicorhynchus spp. | Lumping of closely related species |
| Morulius chrysophekadion | Morulius chrysophekadion/spp. | Labeo chrysophekadion/spp. | Valid name (reference FishBase 2004) |
| Mystus nemurus | Hemibagrus nemurus | Hemibagrus nemurus | H. nemurus = valid name (reference FishBase 2004) |
| Mystus spp. | Hemibagrus wycki/wyckioides/+ Mystus multiradius/ mysticetus/ singaringan/spp. | Mystus spp. non nemurus | Lumping of closely related species, except H. nemurus |
| Pangasius hypophthalmus/sp. | Pangasianodon hypophthalmus | Pangasius hypophthalmus | Pangasius = valid genus (reference FishBase 2004) |
| Parambassis wolffi | Parambassis wolffi/spp. | Parambassis spp. |  |
| Polynemus multifilis | Polynemus longipectoralis (valid name $=$ P. dubius) | Polynemus sp. | Dispute over identification |
| Probarbus jullieni | Probarbus jullieni/ labeamajor | Probarbus spp. | Lumping of two closely related species |
| Puntioplites proctozysron | Puntioplites falcifer | Puntioplites spp. | Dispute over identification |
| Xenentodon <br> sp./ Dermogenys sp. | Xenentodon cancila | Xenentodon cancila | Dermogenys (pusilla, only species) is too small to be caught |

## 6 <br> TEMPORAL ABUNDANCE PATTERNS OF 110 FISH TAXA

Several authors have described the migration patterns of Mekong fishes (review in Baran et al. 2001b). In the past few years a considerable amount of progress has been made in the knowledge of Mekong River fish migrations. Among the most comprehensive publications on regional migrations are those of Poulsen and Valbo-Jørgensen (2000), Bao et al. (2001) and MRC (2001), all based at least partially on fishers' knowledge. In Cambodia, recent insights provided by Srun Phallavan and Ngor Peng Bun (2000), Chhuon Kim Chhea (2000), Chanh Sokheng (2000); Chanh Sokheng et al. (2000) and Heng Kong (2002) have been useful, and in Laos, Baird et al. (1999a), Baird and Flaherty (2005) and Baird (2005a) have relied heavily on local knowledge provided by fishers.

In Southern Laos, Singhanouvong et al. (1996a and b), Warren et al. (1998), Baird et al. (2001a, 2003, 2004), Hogan et al. (2004) and Baird and Flaherty (2004) have also detailed the migration patterns of small Cyprinidae carps and Pangasiidae catfishes based mainly on quantitative field data.

Here, we detail the seasonal appearance of fishes in catches for 110 taxa in Southern Laos, on the basis of Khone Falls fishery monitoring over six years. So far, 201 fish species, including 196 native species, have been recorded from the mainstream Mekong River just below the Khone Falls, but not all of those are important in fisheries (Baird 2001).
The analysis is based on the relative monthly abundance of species in catches; thus for each month, the fraction or percentage of the total annual catches for a given species is provided.

All fishing methods are indistinctly lumped, as this is considered to be a comprehensive multi-gear migration survey covering a diversity of species, of gear selectivity, and of hydrological fishing conditions (six annual cycles, 32 distinct fishing gears sampled; 666,000 fish caught). Analysis is performed on raw biomasses, as gears are generally chosen by fishers to be as efficient as possible, which implicitly follows at best the abundance of fish in the river. Analysis of CPUE data is not

possible i) because the diversity of gears, which were not all monitored consistently each year; and ii) because of the biases detailed in section 3-2-1).

The appearance of fish in catches often implies migratory behavior. Here, migrations are understood to constitute any kind of long-distance or short-distance systematic movement by a population of fish. The Khone Falls fisheries are dependent on both, as even fish migrating over short distances at certain times of the year become more easily catchable then (e.g. Probarbus jullieni, Boesemania microlepis, Baird et al. 2001b; Baird 2005d). Formally speaking, this study only reflects the abundance variation in Khone Falls catches, and is not a study of fish migrations between different locations. However, the studies cited above have largely demonstrated that migrations are a major feature of Mekong fishes, and field experience (e.g. Roberts and Baird 1995; Baird et al. 2003; 2004) confirms that the Khong fisheries are principally based on fish migratory behavior. In any case, the inference from site-specific catches to migrations is appropriate for most species.
Figure 18 and its 11 graphs detail the temporal distribution of fish species in the Khone Falls fisheries. For improved readability, species have been ordered in groups of ten following their scores in a Correspondence Analysis. Interactive distribution analyses for individual species can be performed on the CD-ROM companion to this document.











Figure 18: Temporal distribution of fish species in the Khone Falls fisheries

## Conclusions:

Four clear patterns can be isolated:
Graph 1: precocious species arriving at the beginning of the new hydrological year (Oct-Nov).
Probarbus labeamajor, Tor tambroides, Chonerhinus nefastus, Probarbus jullieni, Tetraodon spp., Amphotistius laosensis, Pseudomystus siamensis, Polynemus longipectoralis, Rhinogobius spp., Cirrhinus molitorella.

Graph 2 to 5: having a dominant presence during the dry season (peak in January); with some of them exhibiting a small secondary peak at the beginning of the rainy season (May-June).

Mystacoleucus spp., Mekongina erythrospila, Parambassis wolffi/spp., Euryglossa panoides, Achiroides spp., Glyptothorax spp., Channa marulius/spp., Cirrhinus jullieni, Hemipimelodus borneensis, Arius stormi, Mystus singaringan/spp., Cynoglossus microlepis, Cyclocheilichthys spp., Botia modesta, Garra fasciacauda, Mystus multiradius/mysticetus/spp., Paralaubuca
typus, Bangana behri, Boesemania microlepis, Crossocheilus siamensis, Botia spp, Catlocarpio siamensis, Toxotes microlepis, Thynnichthys thynnoides, Crossocheilus reticulatus, Scaphognathops stejnegeri, Kryptopterus spp., Rasbora spp., Gyrinocheilus pennocki, Pangasius polyuranodon, Bagarius yarrelli/spp., Acantopsis sp. or spp., Sikukia gudgeri, Luciosoma bleekeri, Onychostoma cf. elongatum, Labiobarbus leptocheilus, Hemibagrus wycki, Barbodes altus, Labeo erythropterus, Aaptosyax grypus, Coius undecimradiatus.

## Graph 6 to 8:

species having two equivalent abundance periods (dry and wet season respectively) or being regularly distributed all year long.

Clarias batrachus/spp., Chitala ornata, Bagrichthys spp., Henicorhynchus spp., Poropuntius deauratus, Macrognathus siamensis/spp., Hypsibarbus malcolmi, Chitala blanci, Cirrhinus microlepis, Osteochilus microcephalus/spp., Glossogobius koragensis, Cosmochilus harmandi, Micronema apogon - micronema, Systomus orphoides, Amblyrhynchichthys truncatus, Channa striata, Notopterus notopterus, Laides hexanema/spp., Raiamas guttatus, Pangasius sanitwongsei, Ompok bimaculatus, Lobocheilos melanotaenia, Hypsibarbus lagleri, Mastacemblus armatus/spp., Scaphognathops bandanensis, Osteochilus spp., Hemisilurus mekongensis, Osteochilus melanopleurus, Helicophagus waandersi, Pangasius pleurotaenia.

Graph 9 to 11: $\quad$ species showing a dominant or exclusive peak at the beginning of the rainy season (May-June).

Labeo chrysophekadion/spp., Hampala macrolepidota, Leptobarbus hoeveni, Belodontichthys dinema, Opsarius spp., Hemibagrus nemurus, Hemibagrus wyckioides, Wallago leeri, Puntioplites falcifer, Hypsibarbus wetmorei, Pristolepis fasciata, Hypsibarbus pierrei, Cirrhinus mrigala, Xenentodon cancila, Cyclocheilichthys enoplos, Osphronemus exodon, Pangasianodon hypophthalmus, Tenualosa thibaudeaui, Pangasius bocourti, Hampala dispar, Setipinna melanochir, Micronema bleekeri, Macrochirichthys macrochirus, Wallago attu, Pangasius larnaudiei, Pangasius conchophilus, Pangasius macronema, Cyprinus carpio, Pangasius krempfi.

These patterns are in general agreement with the interviews of fishers along the Mekong River conducted by the Mekong River Commission (MRC), although there are some differences (Bouakhamvongsa and Poulsen 2000). These patterns can largely be explained by the fact that a large number of species migrate from tributaries into the mainstream Mekong River at the end of the rainy season, when water levels drop. They enter the tributaries again when water levels rise during the rainy season. Moreover, fish are generally easier to catch when water levels are low, and


Figure 19: Major migration patterns in the Mekong mainstream, after Bouakhamvongsea and Poulsen (2000) fishers often have more free time for fishing during the dry season.

Furthermore, a detailed comparison of the above results with the MRC "Fish migrations and spawning habits" survey (Poulsen and Valbo-Jørgensen 2001) gives the following results for 28 species considered in both studies:

- 18 species having essentially the same results
- 7 cases with minimal variation:

| Migration <br> pattern | Poulsen and Valbo-Jorgensen <br> $(2001)$ | Present <br> study |
| :--- | :--- | :--- |
| Paralaubuca typus | Two migrations per year | Second migration non-visible |
| Pangasius polyuranodon | Onset of the flood season | Unclear pattern |
| Aaptosyax grypus | No mention of a second migration | Second migration in June |
| Cirrhinus microlepis | Unclear patterns | Peak in January |
| Hypsibarbus malcolmi | Peak in June | Peak in May |
| Wallago attu | Peak in July | Peak in June |
| Pangasius krempfi |  |  |

- 3 disagreements:

| Migration <br> pattern | Poulsen and Valbo-Jorgensen <br> (2001) | Present <br> study |
| :--- | :--- | :--- |
| Mekongina erythrospilla | March-May | Clear peak in January |
| Botia modesta | Peak in June-July in Khong | Peak in February |
| Pangasius hypophthalmus | Nov.-Dec. peak in Khong | Peak in May-June |

The two studies compared here are highly complementary, as:

- The present study is based on actual catch data over a six-year period, while the study reported on by Poulsen and Valbo-Jørgensen (2001) is based on interviews of fishers (the drawbacks of this latter method are detailed in Poizat and Baran 1997).
- The present study provides the details of migration patterns over all the months of the year and above all encompasses 110 taxa, when that of Poulsen and Valbo-Jørgensen (2001) provides information for only 50 species. The latter study failed to study some of the most important fish species, including Scaphognathops bandenensis.
- As opposed to Poulsen and Valbo-Jørgensen (2001), the present study does not provide any indication of the direction, amplitude or purpose of migrations, but Baird et al. (1999a) does provide much information of this nature for all the species recorded in the Khone Falls fisheries database, as does Baird et al. (2001a; 2003; 2004), and Roberts and Baird (1995).


## HYDROLOGICAL TRIGGERS OF MIGRATIONS: THE EXAMPLE OF THE CATFISH PANGASIUS KREMPFI

Various factors are known to trigger fish migrations, including rises in water levels and changes in discharge, and this study confirms the relationship between migrations and hydrological changes. The catfish Pangasius krempfi provides a remarkable example of the sensitivity of a highly migratory species to water level rises, and warns us of how changes in water levels caused by large dam construction could negatively impact fish populations through changing hydrological conditions necessary for triggering migrations (see, also, Hogan et al. 2005).

The analysis of the daily catches of Pangasius krempfi in 12-16 cm gill nets vs. water level as recorded in Pakse (Figure 20) shows the close relationship between the migrations of this species and rising water levels. The migration occurs suddenly at Khone, and lasts around 40-50 days. However there are in general a dozen days each year when migrations peak, just when the water starts rising after the driest period of the year (e.g. evolution of the catch from 0 to 270 kg per day within 10 days in 1993). This phenomenon is well known by fishers who believe that migrations of P. krempfi are triggered by the first seasonal increases in current speeds. Graphs also show that should the water recede after a first rise, the migration correspondingly slows down. It re-starts as soon as water levels rise again (examples in 1997 and 1998).

These results indicate that dams regulating the downstream hydrology of the Mekong River or one its tributaries would likely have very serious negative impacts on the behavior and migrations of this species, as well as on other species responding similarly to hydrological triggers. One serious example of a large dam that has had a massive impact on the downstream hydrology of a large river is the Yali Falls dam on the Sesan River in the Central Highlands of Viet Nam, which has affected downstream hydrology for hundreds of kilometres in Cambodia (Baird and Dearden 2003).


The statistical quantification of the relationships between fish abundance and water levels is not detailed here because:

1) The relationship is graphically clear and does not require further testing;
2) The day-to-day variability of catches is a major impediment to calculating the correlation between catches (which vary a lot) and water levels (which vary a little), unless catch data is normalized by a disputable $\operatorname{Ln}(\mathrm{x}+1)$ transformation;
3) The catches on day $D$ might depend on water levels on day $D-n$, but $n$ remains undetermined;
4) Beyond that particular case it can be considered that the rise of water levels is a trigger that initiates migrations, but that once initiated, the migrations occur relatively independently from water levels. In that sense these would be events of stochastic nature for which a classical statistical approach based on correlations would be irrelevant.

The impossibility of addressing the migration triggers with standard statistical methods led to prefer a graphic approach, as detailed in the following section.


Figure 20: Relationships between water level increases and Pangasius krempfi catches in Khone Falls fisheries


## 8 <br> DOMINANT SPECIES: LIFE HISTORY KEY FACTS, MIGRATIONS AND TRIGGERS

This section combines biological information originating i) from the Khone Falls fisheries database and ii) from FishBase with hydrological data from the MRC. The objective is to provide an overview of the ecological traits of the dominant species, including major life history key facts (e.g. maximum length, size at maturity, approximate life span, main food, etc), seasonal abundance patterns, in particular in relation to river discharge, and sensitivity to flow modifications.

## Method

## Dominant species

This analysis has been performed on taxa for which more than 750 individuals were caught in Khong between 1993 and 1999 (i.e. at least 50 individuals per year on average); this corresponds to 47 taxa. When only the genus was known, all possible species of this genus present in Laos according to FishBase 2004 have been listed. The original taxonomy used comes from the Khone Falls database, but updated valid names according to FishBase 2004 are also given.

## Life history facts

FishBase gives life history facts for each species. Here we used a new procedure developed upon our request by the FishBase team to automatically produce a "Species ecology matrix", i.e. a table of all ecological key facts for a given list of species (matrix Species x Key facts). This matrix combines original values from the literature (in particular maximum length) and values calculated from other existing parameters (details are given below). In the Khone Falls fisheries database,

when fishes were not known at the species level but only at the genus level, we gave the range of values found for the Lao species of the genus.

Maximum length: The default value used here is the maximum length ever reported in the literature for the species in question.

Approximate life span: This is the maximum age that a fish of the species would reach. It is calculated from the length at infinity, that is itself calculated from growth studies, or when no growth studies are available, from maximum length using an empirical relationship between length at infinity and maximum length (Froese and Binohlan 2000).

Length for maximum yield: This is the size of fish that would result in the maximum possible yield given their size and abundance. This parameter is estimated from growth studies or from length at infinity The matrix of species and corresponding life history key facts is detailed in Annex A.

Main food items: from the literature reviewed in FishBase.
This matrix is supplemented, for each taxon, by a brief description of the seasonal abundance pattern based on figures from section III-3.

## Abundance patterns and discharge

Daily overall catches have been calculated for each taxon, based on the sum of all gear-specific data. Daily water discharge data in Pakse originating from the MRC have been juxtaposed to daily catch data. Graphs showing abundance vs. discharge have been generated for each taxon (with details for species of a given genus whenever available); these graphs are detailed in Annex B. Based on these graphs, the sensitivity to discharge of 47 taxa to hydrological triggers has been coded as nil, medium or high.

## Results

## 1) Dominant taxa in catches

The forty-seven taxa for which at least 50 individuals have been caught on average per year over six years are:

Bagarius spp.; Bagrichthys spp.; Bangana behri; Barbodes altus; Botia modesta; Botia spp; Cirrhinus microlepis; Cosmochilus harmandi; Crossocheilus reticulatus; Crossocheilus siamensis; Cyclocheilichthys enoplos; Garra fasciacauda; Glyptothorax spp.; Gyrinocheilus pennocki; Hemibagrus nemurus; Hemipimelodus borneensis; Hemisilurus mekongensis; Henicorhynchus spp.; Hypsibarbus malcolmi; Kryptopterus spp.; Labeo erythropterus; Labiobarbus leptocheilus; Laides hexanema/spp.; Lobocheilos melanotaenia; Mekongina erythrospila; Labeo chrysophekadion/spp.; Mystacoleucus spp.; Opsarius spp.; Osteochilus spp.; Pangasius bocourti; Pangasius conchophilus; Pangasius krempfi; Pangasius larnaudii; Pangasius macronema; Pangasius pleurotaenia; Pangasius polyuranodon; Paralaubuca typus; Parambassis wolffi/spp.; Pristolepis fasciata; Probarbus jullieni; Pseudomystus siamensis; Puntioplites falcifer; Rasbora spp.; Scaphognathops bandanensis; Sikukia gudgeri; Tenualosa thibaudeaui; Thynnichthys thynnoides.

## 2) Length of fish caught

For each of the dominant taxa, the FishBase matrix gives the maximal length known; this consists of a unique value for each species, or a range of values for a genus. This analysis highlights the kind of fish caught (either large or small species), which is an indicator of the health of the fishery: the higher the percentage of large species, the healthier the fishery is assumed to be.


Figure 21: Maximum size per taxon

The distribution of maximum sizes per taxon is given in Figure 21.
The number of taxa and the biomass caught per class of maximum size are expressed below in terms of relative frequencies (this analysis is based on 43 species sensu stricto for which the maximum length range is well defined):


Figure 22: Maximum size distribution among the dominant taxa


Figure 23: Maximum size distributionin the total biomass caught

These figures show that:

- $61 \%$ of dominant taxa caught have a maximum size superior to 25 cm , but
- $85 \%$ of fishes actually caught (in terms of biomass) belong to taxa whose maximum sizes are inferior to 25 cm .

This illustrates the abundance of small species in the catch, despite the diversity of a fish community in which large species are largely represented.

## 3) Species migrations

Each of the 47 dominant taxa is characterized by its relative abundance during a year (see section 3). This abundance pattern has been coded according to five possible classes:
i) caught all year long; ii) dominant in dry season; iii) dominant between the rainy and dry seasons; iv) dominant in rainy season; v) peaks during both dry and in rainy seasons.

The figure below shows the relative distribution of dominant taxa in terms of temporal abundance.


Figure 24: Temporal abundance patterns of 47 dominant species

These figures lead to the conclusion that $96 \%$ of dominant taxa are not present all year long, and thus undertake migrations or become impossible to catch at certain periods of time (selectivity and operation mode of a gear depending upon hydrology). Given the diversity of fishing methods in the Khone Falls area (see section 2 and 3) aimed at catching fishes whenever they are present, this figure mostly highlights the importance of migrating behavior among dominant species of the Khone Falls fisheries.

## 4) Species, discharge and migration triggers

The relationship between discharge and abundance is illustrated for each of the 47 dominant taxa in Annex B. Figure 25, resulting from an analysis based on all catches over six years and including all the 110 taxa, shows clearly that the highest biodiversity appears in catches during times with the lowest discharge levels. This underscores the importance of dry season water levels for fish and fishers, as well as the importance of rising waters at the end of the dry season. These rising water levels actually act as triggers for several taxa, as illustrated by Pangasius krempfi in section III-4.


Figure 25: Number of taxa caught versus Discharge

The sensitivity of taxa, to discharge variations has been analysed further with the 47 dominant taxa of the Khone Falls database. For each taxon, the sensitivity to discharge has been coded; this sensitivity is expressed by the nature of the relationship between abundance and discharge in the figures of Appendix B, and is coded as nil, medium, high or unknown. The abundance of each taxon in overall catches has also been integrated.


Figure 26: Percentage of each class of sensitivity to discharge among the dominant taxa.


Figure 27: Percentage of taxa sensitive to discharge in the total biomass caught

These results show:

1) The high percentage of taxa sensitive to discharge among the dominant taxa of the Khone Falls
2) The extremely high percentage of species highly sensitive to discharge in the overall catch of the Khone Falls fisheries.

As a conclusion, one can note that when flows vary naturally, the diversity of specific responses to flooding makes some hydrological years good for certain fish species and bad for others; overall a certain catch, made of variable species, is likely year after year. But when flows are regularized, a range of species are permanently impacted and disappear from the catch, which in the long term results in an overall loss. This loss can be very significant when the river concerned is home to the most intensive inland fishery worldwide.


## 9 <br> DEEP-WATER POOLS AS FISH REFUGES

The importance of deep-water pools in the mainstream to the ecology of Mekong fishes has been outlined in several studies or publications (Blanc 1959, Nguyen Xuan Tan and Nguyen Van Hao 1991, Roberts and Baird 1995). Baird et al. (1999a), Chea Vannaren and Sien Kin (2000), Baird et al. (2001a and b), Poulsen and Valbo-Jørgensen (2001), Poulsen et al. (2002), Kolding et al. (2002), Baird and Flaherty (2005) and Baird (2005a and c) have recently addressed this topic again.

It appears that deep pools shelter many species during the dry season; they are also thought to be spawning areas for several species (see, for example, Baird et al. 2001b; Baird and Flaherty 2005). Poulsen et al. (2002) give a list of 39 species reported to use deep pools as a dry season habitat; and Baird (2005c) wrote that fishers from mainstream Mekong River villages in Khong District (Champasak Province, Southern Laos) reported that 51 species had directly benefited from the establishment of Fish Conservation Zones for protecting deep-water pools from being fished. However, beyond fishers' reports, there is no biological data to confirm these hypotheses in the Mekong Basin, although Kolding et al. (2002) reported that hydro-acoustic methods piloted in Khong District have indicated that deep-water pools in the mainstream Mekong River are in fact inhabited by large numbers of fish, albeit, without being able to identify species.

In part of the mainstream Mekong River in Khong District, several deep-water pools have been identified (Baird et al. 1999a and b; Baird et al. 2001b, Baird and Flaherty 2005, Baird 2005a and c), and more specifically, some important deep-water water pools have been monitored in front of Hang Khone Village, Khong District, just below the Khone Falls, right on the border with Northeastern Cambodia (Roberts and Baird, 1995; Baird et al., 2001b; Baird 2005c).


Our objective in this study is to compare the catches of the same gear set either at the surface or in relatively shallow areas outside of deep-water pools or near the bottom of deep-water pools (approximately 20 m deep) in front of Hang Khone Village, and to assess if there is a significant difference between in CPUEs and catch compositions for both fisheries.

The gears studied are monofilament gill nets with mesh sizes ranging from 4 to 9 cm (see section 2 for details about this gear). Surface gill nets were monitored, as a dominant gear, every year from 1993 to 1998, but deep-set gill nets were monitored only in 1994, 1997 and 1998. Surface gill nets are used all year long (with a peak between December and February) while deep-set gill nets are used almost exclusively in February and March ( $98 \%$ of annual use), as these are the dry season months during which fish concentrate in deep pools.

Subsequently bottom and surface catches were compared in February and March 1994, 1997 and 1998. During these months 150 deep fishing operations and 218 surface fishing operations were monitored, time spent fishing was recorded, and 2153 fish belonging to 68 species were caught. The exact surface of each individual gill net was unknown, so the fishing unit was "one net", under the assumption (confirmed by field observations) that the average surface areas of depth gill nets are very similar to those of surface gill nets.

## 9-1) Comparison between deep-water pools and surface CPUEs

Deep-water and surface $4-9 \mathrm{~cm}$ meshed gill nets are compared below during the same period and in the same general areas, in order to determine whether the CPUE of gill nets used in deep-water pools is different from those of the same gear used near the surface or in relatively shallow areas. The CPUE for surface and depth gill nets are given in Table 7.

Table 7: Catches per unit effort (CPUE) of surface and deep gill nets at Hang Khone village

|  | Hours spent fishing |  | Catch (kg) |  | CPUE <br> (grams per hour per net) |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Depth gill <br> net set 4-9 <br> (gm | Surface gill <br> net set 4-9mm | Depth gill net <br> set 4-9 mm | Surface gill <br> net set 4-9 mm | Depth gill <br> net set 4-9 mm | Surface gill <br> net set 4-9 mm |
| March $\mathbf{1 9 9 4}$ | 1284 | 660 | 193.7 | 7.9 | $\mathbf{1 5 1}$ | $\mathbf{1 2}$ |
| February $\mathbf{1 9 9 7}$ | 12 | 1665 | 2.5 | 60.4 | $\mathbf{2 0 6}$ | $\mathbf{3 6}$ |
| February $\mathbf{1 9 9 8}$ | 504 | 948 | 66.4 | 34.0 | $\mathbf{1 3 2}$ | $\mathbf{3 6}$ |

This analysis shows that the CPUEs of gill nets set in deep-water pools in February and March are three to twelve times higher than the CPUEs of surface nets during the same months (see, also, Baird 2005c). This demonstrates the concentration of fishes in deep-water pools during the dry season. Fishers are well aware of this high concentration of fishes, but their fishing effort is tempered by the frequent loss of gears and reduced lifespan of nylon gill nets in rocky deep-water areas.


Figure 28: Biomass in surface and pools of the $\mathbf{3 0}$ most abundant species

## 9-2) Deep-water pool species versus surface species

The following analysis, based on the same data set used above, focuses on the species composition in the catches of $4-9 \mathrm{~cm}$ meshed deep-water and surface gill nets.

The catch of each species in February and March has been divided by the unit effort, and species ordered according to their decreasing abundance. Results are presented in Figure 28 for the thirty dominant species.

The most abundant species in catches (number of individuals caught) are Gyrinocheilus pennocki (a suctorial algae feeder), Hemipimelodus borneensis (an Aridae catfish), Cosmochilus harmandi (a relatively common upper Mekong Cyprinid), Pangasius conchophilus (a large catfish reaching 1.2 m ), and Mekongina erythrospila (a Mekong endemic Cyprinid). However in terms of total biomass, the dominant species are Pangasius conchophilus, Gyrinocheilus pennocki, Hemisilurus mekongensis (a Mekong endemic catfish), Cosmochilus harmandi and Morulius spp. (or Labeo spp., including Labeo chrysophekadion, a large Cyprinid). This latter list gives prominence to species abundant but also of large sizes.

For all these species, CPUE is higher in deep pools than at the surface, with the only exceptions being Mekongina erythrospila and Labeo erythropterus, for which CPUE is higher at the surface.

In order to make these results more readable and integrative, we calculated for each of the 30 species the difference between abundance in depth and at the surface, as well as the difference between biomass in depth and at the surface. Then a bi-plot of species was created (Figure 29).


Figure 29: Dominance of biomass and abundance in each habitat, for the 30 dominant species

This analysis shows that:

- Gyrinocheilus pennocki, Hemipimelodus borneensis, Cosmochilus harmandi, Pangasius conchophilus and Mekongina erythrospila are the five dominant species in the $4-9 \mathrm{~mm}$ gill net peak dry season fishery; they make up to $46 \%$ of the catch (with $16 \%$ for Gyrinocheilus pennocki alone).
- Ten species among the thirty dominant species caught during this monitoring express a clear preference for the bottoms of deep-water pools in the dry season; they are Pangasius conchophilus, Hemisilurus mekongensis, Hemipimelodus borneensis, Hemibagrus wycki, Gyrinocheilus pennocki; Labeo chrysophekadion/spp; Cosmochilus harmandi; Arius stormii; Bagrichthys spp. and Pangasius bocourti.
- At the bottom of deep-water pools Pangasius conchophilus, Hemibagrus wycki or Labeo chrysophekadion spp. are dominant in terms of biomass, whereas Gyrinocheilus pennocki and Pangasius bocourti are dominant in terms of abundance.
- Mekongina erythrospila and Labeo erythropterus are among the few species dominant at the surface;

Some species listed here have not been identified by Poulsen et al. (2002) as making use of deep pools: Hemisilurus mekongensis, Hemipimelodus borneensis, Hemibagrus wyckii, Gyrinocheilus pennocki, Cosmochilus harmandi, Arius stormii, and Bagrichthys spp. Two of these species, Hemipimelodus borneensis and Arius stormi, are Ariidae catfishes not found above the Khone Falls (Roberts, 1993; Roberts and Baird 1995). Baird and Flaherty (2005) wrote that fishers in Khong reported Hemisilurus mekongensis, Hemibagrus wyckii, and Cosmocheilus harmandi as inhabitants of certain types of deep-water habitats. In fact, G. pennocki are found in rocky areas, and are generally not found in the deepest areas but at mid-depth. It is possible that G. pennocki was particularly caught because there are steep cliffs in the fishing area, and these algae eating fish are probably staying inside holes and around the edges of the rocks adjacent to the deep areas.

It is important, however, to recognize that deep-water pools in the mainstream Mekong River certainly do not represent a single habitat, but rather a multitude of habitats that support different communities of fish, even in a relatively restricted area, such as throughout Khong District. Therefore, studies in other deep-water pools, even within Khong District, and especially above the Khone Falls, will certainly result in different catch compositions (Baird and Flaherty 2005, Baird 2005c).

## 9-3) Size of fishes in deep-water pools

Fishers claim that deep-water pools are places where large fishes are often caught, especially in the dry season; it is assumed that these "large fishes" (large individuals of common species) are old adults that contribute significantly to spawning, hence the importance of protecting deep-water pools from overfishing (Baird et al. 1999a; Baird and Flaherty 2005; Baird 2005c). Below, we analyze the differences in the weights of individuals caught in deep-water pools and at the surface.

For this analysis, the same dataset as above was used, but we selected 18 species, for which at least 10 individuals were caught, and at least 5 individuals were caught by each fishing method (deep-water and surface gill nets). This analysis was performed on gill nets with the same mesh sizes, and they therefore had the same selectivity (no sampling bias). However, the use of a restricted mesh size range (here, 4 to 9 cm ) cannot be considered as suitable for length frequency or cohort analysis, since the smallest, and in particular the largest, individuals cannot be caught. The comparison of average individual weight of fish at the surface and in the bottom of deep pools is given in Table 8.

Table 8: Comparison of average individual weight in surface and in deep-water pools

| Species name | Sample: nb in pools vs. nb in surface | \% heavier in pools |  |
| :---: | :---: | :---: | :---: |
| Parambassis wolffi/spp. | 9/11 | 78 |  |
| Amblyrhynchichthys truncatus | 8/8 | 63 | Individuals larger in pools |
| Pangasius conchophilus | 147/11 | 53 |  |
| Chitala blanci | 5/5 | 50 |  |
| Scaphognathops bandanensis | 29/66 | 48 |  |
| Hemibagrus nemurus | 26/14 | 48 |  |
| Hemipimelodus borneensis | 167/6 | 43 |  |
| Mekongina erythrospila | 7/144 | 42 |  |
| Cosmochilus harmandi | 87/75 | 42 |  |
| Puntioplites falcifer | 49/32 | 41 |  |
| Cyclocheilichthys enoplos | 15/9 | 27 |  |
| Euryglossa panoides | 32/5 | 0 |  |
| Notopterus notopterus | 14/8 | -4 |  |
| Hypsibarbus malcolmi | 20/32 | -5 |  |
| Labeo chrysophekadion/spp. | 48/21 | -7 |  |
| Polynemus longipectoralis | 45/11 | -8 |  |
| Micronema apogon - micronema | 14/14 | -16 | Individuals larger at surface |
| Gyrinocheilus pennocki | 256/89 | -25 |  |

Thus, among the 18 species caught at least five times in each environment:

- For eleven species the individuals caught in deep-water pools are 27 to $78 \%$ larger than those caught at the surface; these species are Parambassis wolffi/spp.; Amblyrhynchichthys truncatus; Pangasius conchophilus, Chilata blanci; Scaphognathops bandanensis; Hemibagrus nemurus; Hemipimelodus borneensis; Mekongina erythrospila; Cosmochilus harmandi; Puntioplites falcifer; and Cyclocheilichthys enoplos.
- For two species (Micronema apogon/micronema and Gyrinocheilus pennocki) the individuals caught at the surface were larger than those caught in deep-water pools. However, it is important to recognize that only juvenile Micronema spp. appeared in those results due to the relatively small mesh-size of the gill nets used. It is very rare to catch large individuals of these species in shallow waters, and adults are widely associated with deep-water pools (Baird et al. 1999a and b). No analysis could be made on the reproductive status of these species.

This brief analysis of fish catches at the surface and at the bottom of deep-water pools confirms the important role of these habitats as refuges in the dry season. During those months:

- CPUEs from the bottom of deep-water pools are three to twelve times higher than at the surface.
- Species exhibit various degrees of preference for deep-water pools; however among 30 dominant species, ten are clearly more abundant at the bottom, while two only are more abundant at the surface.
- Within a species, among 18 species studied, individuals of 11 are much larger ( 27 to $78 \%$ larger) at the bottom, while only two have larger individuals at the surface. This indirectly confirms the hypothesis according to which deep-water pools play a refuge role for large old individuals, i.e. breeders contributing largely to stock replenishment.
- Overall these results highlight the importance of developing measures so that Mekong deep-water pools become actual freshwater protected areas. It also shows the reason why fishers living along the mainstream Mekong River in Khong District, Champasak Province (both above and below the Khone Falls), have created a large number of Fish Conservation Zones (FCZs) in deep-water pools. Fishers living along the Mekong River in Southern Laos have long recognized the ecological importance of deep-water pools to fish, and have used this knowledge to both assist them in fishing and to protect fisheries resources (see, also, Baird et al. 2001b, Baird and Flaherty 2005, Baird 2005c).




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| ANNEX 1: DOMINANT SPECIES, LIFE HISTORY KEYFACTS, ABUNDANCE PATTERNS AND RESPONSETO DISCHARGE |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Taxon in the Khone Falls fisheries database | \% of total catch | List of possible species and valid names | Family | Max. length (cm) | Approx. life span (years) | Length for max.yield (cm) | Main food | Seasonal abundance patterns | Abundance in relation to discharge | $\begin{aligned} & \text { Sensitivity } \\ & \text { to } \\ & \text { discharge } \end{aligned}$ |
| Bagarius spp. | 0.06 | Bagarius bagarius | Sisoridae | 70-200 | 2.2-8.3 | 46.1-135.3 | mainly animals | Abundant in Oct.-Dec. + secondary peak in rainy season (May-June) | Species caught between 1,000 and $13,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, rather around $5,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | Medium |
|  |  | Bagarius rutilus |  |  |  |  |  |  |  |  |
|  |  | Bagarius suchus |  |  |  |  | mainly animals |  |  |  |
|  |  | Bagarius yarrelli |  |  |  |  | mainly animals |  |  |  |
| Bagrichthys spp. | 0.07 | Bagrichthys macracanthus | Bagridae | 24.9-30 | 6-6.2 | 16-19.4 | plants/detritus + animals | Dry season + rainy season peaks | Sharp and dense peak between 1,500 and $3,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
|  |  | Bagrichthys macropterus |  |  |  |  | zoobenthos mainly animals |  |  |  |
|  |  | Bagrichthys obscurus |  |  |  |  |  |  |  |  |
| Bangana behri | 0.06 | Bangana behri | Cyprinidae | 60 | 20.4 | 39.4 | mainly plants/detritus | Dry season peak + small secondary peak at first rains | Peak between 2,000 and $4,500 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
| Barbodes altus | 0.17 | Barbonymus altus | Cyprinidae | 20 | 8.8 | 12.8 | plants mainly plants/detritus | Dry season peak + small secondary peak at first rains | Species caught between 2,000 and $8,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, with a peak between 2,000 and $3,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
| Botia spp. | 0.17 | Botia caudipunctata | Cobitidae | 5.5-30 | 3.2-15.7 | 3.4-19.4 |  | Dry season peak + small secondary peak at first rains | Very sharp and dense peak between 2,000 and $3,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
|  |  | Botia eos |  |  |  |  | mainly animals |  |  |  |
|  |  | Botia helodes |  |  |  |  | zoobenthos mainly animals |  |  |  |
|  |  | Botia lecontei |  |  |  |  | zoobenthos mainly animals |  |  |  |
|  |  | Botia longidorsalis |  |  |  |  |  |  |  |  |
|  |  | Botia longiventralis |  |  |  |  | plants plants/ detritus + animals |  |  |  |
|  |  | Botia morleti |  |  |  |  | mainly animals |  |  |  |
|  |  | Botia nigrolineata |  |  |  |  |  |  |  |  |
|  |  | Botia sidthimunki |  |  |  |  | zoobenthos mainly animals |  |  |  |
|  |  | Botia splendida |  |  |  |  |  |  |  |  |


| Taxon in the Khone Falls fisheries database | \% of total catch | List of possible species and valid names | Family | Max. length (cm) | Approx. life span (years) | Length for max.yield (cm) | Main food | Seasonal abundance patterns | Abundance in relation to discharge | $\begin{aligned} & \text { Sensitivity } \\ & \text { to } \\ & \text { discharge } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Botia modesta | 14.55 | Botia modesta | Cobitidae | 25 | 13.5 | 16.1 | zoobenthos mainly animals | Dry season peak | Very sharp and dense peak between 2,000 and 3,000 $\mathrm{m}^{3} . \mathrm{s}^{-1}$ | High |
| Cirrhinus microlepis | 0.13 | Cirrhinus microlepis | Cyprinidae | 65 | 26 | 42.7 | plants mainly plants/detritus | Dry season peak + small secondary peak at first rains | Species caught between 1,500 and $7,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, concentrated between 2,000 and $4,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
| Cosmochilus harmandi | 0.40 | Cosmochilus harmandi | Cyprinidae | 100 | 32 | 66.5 | plants mainly plants/detritus | Two peaks, one at decreasing flows (November-December) and one at first rains (May-June) | Species caught between 1,500 and $13,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, concentrated between 1,500 and $6,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | Medium |
| Crossocheilus reticulatus | 3.47 | Crossocheilus reticulatus | Cyprinidae | 17 | 7.6 | 10.8 | plants/detritus + animals | Dry season peak | Species caught between 1,500 and $19,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, with a sharp peak between 1,500 and $2,500 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
| Crossocheilus siamensis | 0.07 | Crossocheilus siamensis | Cyprinidae | 16 | 7.2 | 10.1 | mainly plants/detritus | Dry season peak | Species caught between 1,500 and $10,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, with a sharp and intense peak between 1,500 and $2,500 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ (but 10 data points only) | High |
| Cyclocheilichthys enoplos | 0.10 | Cyclocheilichthys enoplos | Cyprinidae | 74 | 28.7 | 48.8 | zooplankton mainly animals | Dominant peak at first rains, but also caught in the dry season | Species caught between 1,500 and $20,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, with a sharp peak around $3,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
| Garra fasciacauda | 1.62 | Garra fasciacauda | Cyprinidae | 11 | 5 | 6.9 | plants/detritus + animals | Dry season peak | Sharp peak between 2,000 and 3,000 m³.s-1, but few data points | High |


| Taxon in the Khone Falls fisheries database | \% of total catch | List of possible species and valid names | Family | Max. length (cm) | Approx. life span (years) | Length for max. yield (cm) | Main food | Seasonal abundance patterns | Abundance in relation to discharge | $\begin{aligned} & \text { Sensitivity } \\ & \text { to } \\ & \text { discharge } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Glyptothorax spp. | 0.13 | Glyptothorax fuscus | Sisoridae | 8.8-30 | 2.9-12.4 | 0.6-19.4 | mainly animals | Dry season peak + small secondary peak at first rains | Species caught between 2,000 and 5,000 m³.s-1, but few data points | Unknown |
|  |  | Glyptothorax honghensis |  |  |  |  |  |  |  |  |
|  |  | Glyptothorax interspinalum |  |  |  |  |  |  |  |  |
|  |  | Glyptothorax lampris |  |  |  |  | mainly animals |  |  |  |
|  |  | Glyptothorax laosensis |  |  |  |  | mainly animals |  |  |  |
|  |  | Glyptothorax macromaculatus |  |  |  |  |  |  |  |  |
|  |  | Glyptothorax major |  |  |  |  | mainly animals |  |  |  |
|  |  | Glyptothorax trilineatus |  |  |  |  |  |  |  |  |
|  |  | Glyptothorax zanaensis |  |  |  |  |  |  |  |  |
| Gyrinocheilus pennocki | 0.73 | Gyrinocheilus pennocki | Gyrinocheilidae | 28 | 3.8 | 18 | plants/detritus + animals | Dry season peak + small secondary peak at first rains | Sharp and dense peak between 1,500 and 3,500 m ${ }^{3}$.s- ${ }^{-1}$ | High |
| Hemibagrus nemurus | 0.22 | Hemibagrus nemurus | Bagridae | 65 | 15.1 | 42.7 | zoobenthos mainly animals | Dominant peak at first rains, secondary peak in the dry season | No clear pattern but concentration around $1,500-9,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | Medium |
| Hemipimelodus borneensis | 0.10 | Hemipimelodus borneensis | Ariidae | 30 | 10.1 | 19.4 | plants/detritus + animals | Appears in November, but major peak in the dry season | No clear pattern | Nil |
| Hemisilurus mekongensis | 0.08 | Hemisilurus mekongensis | Siluridae | 80 | 2.6 | 52.9 | plants/detritus + animals | Two peaks, one in dry season, one in wet season | No pattern | Nil |
| Henicorhynchus spp. ${ }^{3}$ | 47.26 | Henicorhynchus cryptopogon | Cyprinidae | 9.5-20 | 4.3-8.8 | 5.9-12.8 | plants mainly plants/detritus | Dry season peak + small secondary peak at first rains | Pattern species-specific | High |
|  |  | Henicorhynchus lineatus |  |  |  |  |  |  |  |  |
|  |  | Henicorhynchus ornatipinnis |  |  |  |  |  |  |  |  |
|  |  | Henicorhynchus siamensis |  |  |  |  | plants mainly plants/detritus |  |  |  |

Complex group of poorly identified species and controversial taxonomy, including also Cirrhinus spp., in particular Henicorhynchus/Cirrhinus lobatus, dominant in catches and coming from Cambodia where it is not recorded.

| Taxon in the Khone Falls fisheries database | \% of total catch | List of possible species and valid names | Family | Max. length (cm) | Approx. life span (years) | Length for max.yield (cm) | Main food | Seasonal abundance patterns | Abundance in relation to discharge | Sensitivity to discharge |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Hypsibarbus malcolmi | 0.37 | Poropuntius malcolmi | Cyprinidae | 50 | 20.4 | 32.7 | mainly animals | 2 equal peaks, in dry season and wet season | Species caught between 1,500 and $11,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, with a peak between 3,000 and 5,000 $\mathrm{m}^{3} . \mathrm{s}^{-1}$ | High |
| Kryptopterus spp. | 0.18 | Kryptopterus apogon | Siluridae | 12-130 | 2.9-7.6 | 7.6-87 | nekton mainly animals | Peak in December + small secondary peak at first rains (May) | No pattern | Nil |
|  |  | Kryptopterus bicirrhis |  |  |  |  | mainly animals |  |  |  |
|  |  | Kryptopterus cheveyi |  |  |  |  | mainly animals |  |  |  |
|  |  | Kryptopterus cryptopterus |  |  |  |  | nekton mainly animals |  |  |  |
|  |  | Kryptopterus dissitus |  |  |  |  |  |  |  |  |
|  |  | Kryptopterus geminus |  |  |  |  |  |  |  |  |
|  |  | Kryptopterus hexapterus |  |  |  |  | mainly animals |  |  |  |
|  |  | Kryptopterus limpok |  |  |  |  | mainly animals |  |  |  |
|  |  | Kryptopterus micronema |  |  |  |  | mainly animals |  |  |  |
|  |  | Kryptopterus moorei |  |  |  |  | mainly animals |  |  |  |
|  |  | Kryptopterus schilbeides |  |  |  |  | mainly animals |  |  |  |
| Labeo erythropterus | 0.09 | Labeo erythropterus | Cyprinidae | 70 | 28.6 | 46.1 | mainly plants/detritus | Peak in December + secondary peak at first rains (June) | Species caught between 1,500 and $33,000 \mathrm{~m}^{3} \cdot \mathrm{~s}^{-1}$, concentrated between 2,000 and $4,500 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
| Labiobarbus leptocheilus | 6.35 | Labiobarbus leptocheila | Cyprinidae | 30 | 12.9 | 19.4 | plants mainly plants/detritus | Dry season dominant peak + small secondary peak in rainy season | Species caught between 1,500 and $21,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, with a concentration between 1,500 and $3,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
| Laides hexanema/spp. | 0.06 | Laides hexanema | Schilbeidae | 14.2-16.5 | 4.5-4.8 | 9-10.5 |  | Almost two equal peaks in dry and early rainy season | Species caught between 2,000 and $5,500 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
|  |  | Laides longibarbis |  |  |  |  | mainly animals |  |  |  |


| Taxon in the Khone Falls fisheries database | \% of total catch | List of possible species and valid names | Family | Max. length (cm) | Approx. life span (years) | Length for max. yield (cm) | Main food | Seasonal abundance patterns | Abundance in relation to discharge | Sensitivity to discharge |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Lobocheilos melanotaenia | 1.25 | Lobocheilos melanotaenia | Cyprinidae | 20 | 8.8 | 12.8 | plants mainly plants/detritus | Dry season peak + small secondary peak in early rainy season | Species caught between 1,500 and $24,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$; no clear pattern | Nil |
| Mekongina erythrospila | 0.57 | Mekongina erythrospila | Cyprinidae | 45 | 19 | 29.3 | mainly plants/detritus | Dry season peak (January) | Very sharp and dense peak between 1,500 and $3,500 \mathrm{~m}^{3} . \mathrm{s}-{ }^{-1}$, in particular around $2,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
| Labeo chrysophekadion/spp. | 0.06 | Labeo chrysophekadion | Cyprinidae | 90 | 28.7 | 59.7 | plants mainly plants/detritus | Dominant peak at first rains, secondary peak in the dry season | Species caught between 1,500 and $26,000 \mathrm{~m}^{3} \cdot \mathrm{~s}^{-1}$, concentrated between 1,500 and $5,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | Medium |
| Mystacoleucus spp. | 0.11 | Mystacoleucus atridorsalis | Cyprinidae | 7.3-20 | 3.4-7.4 | 4.5-12.8 | plants/detritus + animals | Dominant dry season peak + small secondary peak at first rains | Peak between 2,000 and $5,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
|  |  | Mystacoleucus chilopterus |  |  |  |  |  |  |  |  |
|  |  | Mystacoleucus ectypus |  |  |  |  |  |  |  |  |
|  |  | Mystacoleucus greenwayi |  |  |  |  |  |  |  |  |
|  |  | Mystacoleucus lepturus |  |  |  |  |  |  |  |  |
|  |  | Mystacoleucus marginatus |  |  |  |  | plants/detritus + animals |  |  |  |
| Opsarius spp. | 0.06 | Opsarius koratensis | Cyprinidae | 10 | 4.6 | 6.3 | zoobenthos mainly animals | Dominant peak at first rains, secondary peak in the dry season | Insufficient data | Unknown |


| Taxon in the Khone Falls fisheries database | \% of total catch | List of possible species and valid names | Family | Max. length (cm) | Approx. life span (years) | Length for max.yield (cm) | Main food | Seasonal abundance patterns | Abundance in relation to discharge | Sensitivity to discharge |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Osteochilus spp. | 0.08 | Osteochilus enn eaporus | Cyprinidae | 8.2-60 | 2.5-23.8 | 5.1-39.4 | mainly plants/detritus | Caught almost all year long (except in April and Sept.) but speciesspecific variability | Pattern species-specific | Unknown |
|  |  | Osteochilus hasseltii |  |  |  |  | zoobenthos plants/detritus + animals |  |  |  |
|  |  | Osteochilus lini |  |  |  |  | mainly plants/detritus |  |  |  |
|  |  | Osteochilus melanopleurus |  |  |  |  | plants mainly plants/detritus |  |  |  |
|  |  | Osteochilus microcephalus |  |  |  |  | mainly plants/detritus |  |  |  |
|  |  | Osteochilus salsburyi |  |  |  |  |  |  |  |  |
|  |  | Osteochilus striatus |  |  |  |  |  |  |  |  |
|  |  | Osteochilus vittatus |  |  |  |  |  |  |  |  |
|  |  | Osteochilus waandersii |  |  |  |  | mainly plants/detritus |  |  |  |
| Pangasius bocourti | 0.14 | Pangasius bocourti | Pangasiidae | 120 | 17.5 | 80.2 | plants/detritus + animals | Dominant peak in early rainy season, secondary peak in Nov. | Species caught between 2,00 and $31,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$; no clear pattern | Medium |
| Pangasius conchophilus | 0.59 | Pangasius conchophilus | Pangasiidae | 120 | 24 | 80.2 | mainly animals | Exclusive peak in early rainy season | Species caught between 1,500 and 25,000 m${ }^{3} . \mathrm{s}^{-1}$, concentrated between 3,000 and $6,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | Medium |
| Pangasius krempfi | 0.25 | Pangasius krempfi | Pangasiidae | 120 | 2.6 | 80.2 |  | Exclusive peak in early rainy season | Species caught between 1,500 and 20,000 m ${ }^{3} . \mathrm{s}^{-1}$, concentrated between 2,000 and $9,500 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
| Pangasius larnaudiei | 0.07 | Pangasius larnaudii | Pangasiidae | 130 | 2.7 | 87 | zoobenthos mainly animals | Exclusive peak in early rainy season | Species caught between 1,500 and $22,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, concentrated between 2,500 and $10,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | Medium |
| Pangasius macronema | 0.26 | Pangasius macronema | Pangasiidae | 30 | 2.6 | 19.4 | zoobenthos mainly animals | Exclusive peak at first rains | Species caught between 1,500 and $26,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, with a sharp and intense peak between 3,000 and $4,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
| Pangasius pleurotaenia | 0.17 | Pangasius pleurotaenia | Pangasiidae | 35 | 2.4 | 22.7 | plants/detritus + animals | Dominant peak at first rains, secondary peak in the dry season | No pattern | Nil |


| Taxon in the Khone Falls fisheries database | \% of total catch | List of possible species and valid names | Family | Max. length (cm) | Approx. life span (years) | Length for max. yield (cm) | Main food | Seasonal abundance patterns | Abundance in relation to discharge | $\begin{gathered} \text { Sensitivity } \\ \text { to } \\ \text { discharge } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Pangasius polyuranodon ${ }^{4}$ | 0.15 | Pangasius polyuranodon | Pangasiidae | 80 | 80 | 80 | plants/detritus + animals | Present all year long (mainly in Oct.) except in April | No pattern | Nil |
| Paralaubuca typus | 14.63 | Paralaubuca typus | Cyprinidae | 18 | 8 | 11.4 | zooplankton mainly animals | Exclusive dry season peak | Species caught between 1,500 and $5,500 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, with a sharp and intense peak around $2,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
| Parambassis spp. | 0.07 | Parambassis apogonoides | Ambassidae | 6-20 | 1.7-5.3 | 3.7-12.8 | zoobenthos mainly animals | Exclusive dry season peak | Peak between 1,500 and 4,000 $\mathrm{m}^{3} . \mathrm{s}^{-1}$ | High |
|  |  | Parambassis siamensis |  |  |  |  |  |  |  |  |
|  |  | Parambassis wolffii |  |  |  |  | zoobenthos mainly animals |  |  |  |
| Pristolepis fasciata | 0.06 | Pristolepis fasciata | Nandidae | 20 | 4.4 | 12.8 | zoobenthos mainly animals | Dominant peak in the rainy season (July) | No pattern | Nil |
| Probarbus jullieni | 0.13 | Probarbus jullieni | Cyprinidae | 150 | 57.7 | 100.8 | plants/detritus + animals | Dominant peak in Dec. + secondary peak in May | Species caught between 1,500 and $8,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, with a peak between 3,000 and $5,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | Medium |
| Pseudomystus siamensis | 0.10 | Pseudomystus siamensis | Bagridae | 20 | 4.2 | 12.8 | zoobenthos mainly animals | Dominant peak in Dec. + secondary peak in July | No pattern | Nil |
| Puntioplites falcifer | 0.40 | Puntioplites falcifer | Cyprinidae | 40 | 13.6 | 26 | mainly plants/detritus | Dominant peak at first rains, secondary peak in the dry season | Species caught between 1,500 and $7,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, with a concentration between 2,000 and $5,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |


|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
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|  | $\begin{aligned} & \infty \\ & \stackrel{\infty}{6} \\ & \hline \stackrel{1}{0} \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | $\stackrel{\circ}{\stackrel{\circ}{\mathrm{m}}}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | $\stackrel{\hat{\vdots}}{\dot{\mathrm{j}}}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| $\begin{aligned} & \stackrel{\rightharpoonup}{6} \\ & \stackrel{y}{6} \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  | $\left\|\begin{array}{r} n \\ n \\ 0 \\ 0 \\ 0 \\ 0 \\ 0.0 \\ 0.0 \\ 0.0 \\ 0 \\ 0 \\ 0 \\ 0 \\ 0 \end{array}\right\|$ |  | $\left\|\begin{array}{\|c\|} 0 \\ 0 \\ 0 \\ 0 \\ 0 . \\ 0 \\ 0 \\ 0 \\ 0 \\ 0 \\ 0 \\ 0 \\ 0 \\ 0 \end{array}\right\|$ | $n$ 0.4 0 0 0 0 0 0 0 0 0 0 0 0 |  |  |  | $\begin{aligned} & \text { İ } \\ & \text { O} \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 . \\ & 0 \\ & 0 \end{aligned}$ |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  | $\bigcirc$ |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Taxon in the Khone Falls fisheries database | \% of total catch | List of possible species and valid names | Family | Max. length (cm) | Approx. life span (years) | Length for max. yield (cm) | Main food | Seasonal abundance patterns | Abundance in relation to discharge | Sensitivity to discharge |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Scaphognathops bandanensis | 1.65 | Scaphognathops bandanensis | Cyprinidae | 40 | 16.7 | 26 | plants/detritus + animals | 2 equal peaks, in dry season and wet season | Species caught between 1,500 and $21,000 \mathrm{~m}^{3} \cdot \mathrm{~s}^{-1}$, with a peak around $3,500 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
| Sikukia gudgeri | 0.18 | Sikukia gudgeri | Cyprinidae | 18 | 8 | 11.4 | plants/detritus + animals | Almost exclusive peak in the dry season | Species caught between 2,000 and $22,500 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, with a sharp peak between 2,000 and $3,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | Medium |
| Tenualosa thibaudeaui | 0.16 | Tenualosa thibaudeaui | Clupeidae | 30 | 7.3 | 19.4 | plants mainly plants/detritus | Dominant peak at first rains, secondary peak in the dry season | Species caught between 2,000 and $5,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |
| Thynnichthys thynnoides | 1.22 | Thynnichthys thynnoides | Cyprinidae | 25 | 9.1 | 16.1 | plants mainly plants/detritus | Almost exclusive peak in the dry season | Species caught between 2,500 and $12,500 \mathrm{~m}^{3} . \mathrm{s}^{-1}$, with a sharp peak between 2,000 and $3,000 \mathrm{~m}^{3} . \mathrm{s}^{-1}$ | High |

## ANNEX 2: RELATIONSHIP BETWEEN ABUNDANCE AND








## RIVER DISCHARGE FOR DOMINANT SPECIES













Discharge $(1,000 \mathrm{~m} 3 / \mathrm{s})$










Species: Pangasius conchophilus


Species: Pangasius pleurotaenia


Species: Pristolepis fasciata



Species: Pangasius krempfi


Species: Probarbus jullieni











[^0]:    ${ }^{1}$ Henicorhynchus lobatus (Smith 1945) is referenced as Cirrhinus lobatus (Smith 1945) in FishBase 2004, based on Roberts (1997).

[^1]:    ${ }^{1}$ The population and Development Database, http://www.alsagerschool.co.uk/subjects/sub_content/geography/Gpop/HTMLENH/country/kh.htm

